

1 **RESEARCH ARTICLE**

2 **Wildlife mortality on an Ecuadorian coastal highway**

3 Catherine L. Woodward^{1*}, Luis M. Fernández, and Joe E. Meisel

4 *Ceiba Foundation for Tropical Conservation, 301 S. Bedford St, Suite 7A, Madison, WI*
5 *53703, United States; Phone: +1 608 230-5550. Juan Ramírez N36-14 y Germán Alemán,*
6 *170504, Quito, Ecuador; Phone: +593 2603-5904*

7 *29 September 2016*

8 Transportation networks have direct and indirect impacts on the ecosystems they
9 traverse. Roads have become an omnipresent element that modify and fragment
10 ecosystems and divide animal populations. The most direct effect of roads is the death
11 of the animals that try to cross them as they move between habitat patches. We studied
12 wildlife mortality along a 63-km stretch of coastal highway in Manabí, Ecuador. We
13 conducted 12 surveys from April 2014 to July 2015 recording every animal carcass
14 visible within the roadway and its GPS coordinates. We found a total of 916 wild
15 animal carcasses, comprised of 501 amphibians (54.7%), 225 birds (24.6%), 103
16 reptiles (11.2%), and 77 mammals (8.4%). Ten animals (1.1%) could not be identified.
17 Assuming these surveys are representative of mortality rates throughout the year, we
18 estimated that the total number of animals killed on this stretch of road is between
19 1,860 (assuming a 15-day carcass turnover rate (CTR)) and 9,300 (assuming a 3-day
20 CTR). Our results indicate a wide variety of taxa affected, including several threatened
21 species. We identified road mortality “hotspots” in the region, and recommend

¹*Corresponding author. Email: cwoodward@ceiba.org

22 mitigation strategies such as traffics signs and wildlife crossing structures to reduce
23 wildlife mortality in these locations.

24 **Keywords:** roadkill, road ecology, fauna, Ecuador.

25 Las redes de transporte tienen impactos directos e indirectos sobre los ecosistemas que
26 atraviesan. Las carreteras se han convertido en elementos omnipresentes que modifican
27 y fragmentan los ecosistemas y dividen las poblaciones de animales. El efecto más
28 directo de las carreteras es la muerte de los animales que tratan de cruzarlas para
29 moverse entre fragmentos de hábitat. Estudiamos la mortalidad de fauna silvestre a lo
30 largo de 63 km de la autopista costera en Manabí, Ecuador. Realizamos 12 muestreos
31 entre abril de 2014 y julio de 2015 registrando cada cadáver animal encontrado en la
32 vía y sus coordenadas GPS. Encontramos un total de 916 animales silvestres
33 atropellados, conformados por 501 anfibios (54,7%), 225 aves (24,6%), 103 reptiles
34 (11.2%) y 77 mamíferos (8.4%). Diez animales (1,1%) no pudieron ser identificados.
35 Asumiendo que estos muestreos son representativos de las tasas de mortalidad a lo
36 largo del año, estimamos que el total de animales muertos por atropellamientos en este
37 segmento de vía está entre 1860 (asumiendo una tasa de renovación del cadáver (CTR)
38 de 15 días) y 9300 (asumiendo una CTR de 3 días). Nuestros resultados indican que
39 una gran variedad de taxones están afectados, incluyendo varias especies amenazadas.
40 Identificamos "hotspots" de mortalidad en la región y recomendamos varias estrategias
41 de mitigación, tales como señales de tráfico y estructuras de paso de fauna para reducir
42 la mortalidad de la fauna silvestre en estas localizaciones.

43 **Palabras clave:** mortalidad en la vía, ecología de carreteras, fauna, Ecuador.

44 **Introduction**

45 Transportation networks are usually planned based upon the potential economic and social
46 development of the places they connect, and environmental impacts are rarely taken into

47 account unless they cross through natural or protected areas. These networks however, and
48 roads in particular, result in major modifications of the ecosystems they pass through, with
49 animal-vehicle collisions killing millions of animals globally each year [1-3].

50 Impacts of roads on abiotic ecosystem components include altered hydrology and
51 sedimentation, increased noise level, and pollution [3–5]. Roads also have direct and indirect
52 effects on the biotic components of ecosystems [6], including spread of invasive plants [3],
53 and behavioral changes in animals [7], but animal mortality due to vehicular collisions is the
54 main effect of roads on vertebrates [2,3], leading to the reduction of population sizes in
55 amphibians [8], reptiles [9], birds [10], and mammals [11]. This direct impact is especially
56 important in megadiverse countries like Ecuador due to the higher number of affected species
57 and large numbers of rare endemics [5,12].

58 Fragmentation of habitat by roads increases edge effects, reducing the effective
59 habitat area especially for edge-sensitive species [13]. Roads act as a barrier for the
60 movement of wildlife [3, and references therein] thereby reducing genetic exchange and
61 increasing the probability of local extinctions [14,15]. Although road ecology has developed
62 into a discipline in its own right, data from South America are scarce [5]. Existing studies
63 indicate that road mortality is a global threat to wildlife [5,16,17].

64 The main goal of this study is to determine the direct impact of an Ecuadorian coastal
65 highway on terrestrial vertebrate fauna, and provide a baseline for monitoring future wildlife
66 mortality as traffic increases to this developing tourism area. Specifically, we sought to
67 identify the most affected taxa and the most critical wildlife crossing points along the
68 highway in order to make recommendations for mitigating faunal mortality.

69 **Material and methods**

70 We conducted this study along the Pacific coastal highway in Ecuador (E-15) between April
71 2014 and July 2015. This 741 km long highway crosses three out of the five coastal provinces
72 of the country (Esmeraldas, Manabí, and Santa Elena), from the Chocó region in the north to
73 the Tumbes region in the south. We surveyed a 63-km stretch of highway between the towns
74 of Pedernales and Rambuche, in northern Manabí province (Figure 1), that traverses
75 threatened seasonally deciduous tropical forests and transitional ecosystems bridging these
76 two ecoregions. The survey area lies within the Tumbes-Chocó-Magdalena biodiversity
77 hotspot, and two Important Bird Areas (EC010 Hacienda Camarones and EC011 Tito Santos
78 [18]). Although this road has been in place for several years it was paved and widened in
79 2013 [19]. In the area surveyed, the road is two lanes, with an average width of 9 m and an
80 unenforced speed limit of 80 km/h. New housing and tourism developments in the region
81 suggest that traffic will increase along this road in the future.

82 We carried out 12 surveys in a car moving at a constant speed of 30 km/h, with two
83 observers checking both sides of the road, from curb to curb. For each carcass found, we
84 classified it at least to Class whenever possible (e.g., amphibian, reptile, bird, or mammal)
85 and recorded its coordinates (Garmin[®] GPSMap 78). We assumed that all animals found
86 dead in the road were killed by collisions with vehicles. All domestic animals identified,
87 mostly dogs, were not included in the analyses.

88 To estimate the annual roadkill rate (individuals killed/km/year) we used the
89 following formula: $(\text{number of carcasses}/\text{length of road in km}) * (365 \text{ days}/\text{number of days}$
90 $\text{monitored})$. A prior ad hoc study in the region determined that carcasses decompose or are
91 removed or consumed by scavengers in fewer than 15 days (J. Meisel, *unpubl. data*), and
92 published studies have utilized a range of carcass turnover rates (CTR) [e.g., 2,16]. We
93 therefore provide three estimates of annual roadkill reflecting high (CTR = 3 days), medium

94 (CTR = 7 days), and low (CTR = 15 days) carcass turnover rates, that are multiplied by the
95 number of surveys to obtain “*number of days monitored*” in our formula.

96 To identify mortality hotspots, we overlaid each road kill point on a georeferenced
97 map of the road system using QGIS (v 2.14) and ArcMap (v 10.3). Using a spatial join, we
98 obtained the number of kills in each 0.5 km segment of road ($n=126$) for all kills combined
99 as well as for each Class separately, and applied a square-root transformation to the count
100 data to achieve a normal distribution. We defined hotspots as 0.5 km segments where number
101 of kills exceeded two standard deviations (*S.D.*) above the mean (\bar{X}) of the transformed data.
102 To create a map showing mortality hotspots, we assigned a color ramp using five bins
103 separated by values equal to $\bar{X} + 2 S.D.$, $\bar{X} + 1 S.D.$, \bar{X} , and 0 (zero roadkills).

104 **Results**

105 In 12 surveys of this 63 km roadway we observed a total of 916 wild animal carcasses. The
106 most affected taxa were amphibians (501 individuals; 54.7%), followed by birds (225
107 individuals; 24.6%), reptiles (103 individuals; 11.2%), and mammals (77 individuals; 8.4%).
108 There were 10 individuals (1.1%) that were too decomposed or fragmented to be classified
109 (Figure 2). We also encountered 23 domestic animals (21 dogs, one horse, and one chicken)
110 that were excluded from the data analysis.

111 Assuming constant kill rates over the year, the lowest estimate of the annual roadkill
112 rate (CTR = 15 days) is 29.5 roadkills/km, the medium estimate (CTR = 7 days) is 63.2
113 roadkills/km, and the highest estimate (CTR = 3 days) is 147.4 roadkill/km on this stretch of
114 the road per year. These rates translate, respectively, to annual totals of 1857, 3980, and 9287
115 animals killed just on the 63 km length of the road we sampled. If roadkill rates are similar
116 along its entire 741 km length, the number of wild animals killed each year on the Pacific

117 coastal highway of Ecuador could be as high as 109,235. Roadkill rates and annual mortality
118 estimates by taxonomic groups can be found in Table 1.

119 Although most carcasses could be identified to broader taxonomic group (e.g.,
120 opossums), we were able to identify only 90 individuals definitively to genus or species,
121 comprising 30 different taxa. Of them, the most commonly found in each Class were the cane
122 toad (*Rhinella marina*), the boa constrictor (*Boa constrictor*), anis (*Crotophaga* spp.), and
123 the common opossum (*Didelphis marsupialis*). There were several examples of vulnerable
124 species that were identified among the kills, including the Pacific royal flycatcher
125 (*Onychorhynchus occidentalis*), the rufous-headed chachalaca (*Ortalis erythroptera*), and the
126 Northern tamandua (*Tamandua mexicana*) [20, 21].

127 Hotspot mapping identified five 0.5-km segments with combined mortality across all
128 taxa exceeding two standard deviations above the mean of the transformed data (global $\bar{X} =$
129 2.36, $S.D. = 1.30$, $n = 126$); some segments were contiguous, however, resulting in a total
130 of two distinct hotspots (Figure 3A). When amphibians were excluded, the largest of the two
131 hotspots disappeared, but new hotspots were revealed, reflecting road segments with high
132 mortality of other taxa (Figure 3B). Hotspots separated by Class are shown in Figures 3C-F.

133 **Discussion**

134 Our study confirms that roads have a direct harmful effect on wildlife populations across
135 many taxa. Concordant with other studies [22-24], amphibians were the hardest hit,
136 representing over half of all roadkills. Because CTR may vary widely based on the
137 abundance of scavengers, climate and weather conditions, removal by people, and inherent
138 characteristics of the animal bodies, as well as the strongly aggregated pattern of mortality,
139 amphibian roadkill may be most prone to underestimation. Amphibian carcasses are small in
140 size and more likely to be missed by observers, to be removed in one piece by scavengers, or

141 to rapidly decompose [24-26]. In our study, however, we observed toad carcasses (*Rhinella*
142 *marina*) to be extremely persistent, remaining on the road as desiccated crusts for several
143 weeks, particularly in the dry season. Better CTR data are needed on a variety of taxa to
144 improve the reliability of roadkill estimates. Still, we believe the CTR values we used offers a
145 realistic range (21,847 to 109,235) for the number of wild animals killed each year along
146 Ecuador's E-15 coastal highway.

147 The 23 domestic animal carcasses encountered represent just 2.4% of the total
148 roadkills observed, but 25.9% of the identified mammals (22 out of 85). Our total roadkill
149 estimates for wild animals would be slightly overestimated if a similar percentage of the 14
150 mammals in our sample that we could not identify plus the ten carcasses that could not be
151 identified to any taxonomic level were in fact domesticated animals (ca. 6 animals). While it
152 is obviously important to obtain accurate species identifications whenever possible and
153 exclude domestic roadkills to obtain precise estimates, we found the vast majority of animal
154 collisions with cars to involved wild animals.

155 Mapping the distribution of kills through hotspot analysis allowed us to identify
156 critical crossing points where risk of mortality is high. Roadkills occurred throughout the
157 entire length of road surveyed (only four 0.5-km segments had zero mortalities), but because
158 amphibians comprised 54% of the data, they masked mortality hotspots of other taxa.
159 Analysis of hotspots by Class revealed that mammal and reptile mortality were more evenly
160 distributed, while amphibian and bird mortality were more strongly aggregated. There was no
161 overlap between hotspots for amphibians and hotspots for mammals, suggesting that these
162 groups utilize different parts of the landscape.

163 Analysis of the hotspot maps allow us to identify factors that may be predictive of
164 high wildlife mortality. For example, the two hotspots for amphibian mortality are located in

165 low-lying areas, where amphibians may congregate to breed. Hotspots for mammals seem to
166 coincide with areas where forest is close to the road, although further analysis is needed to
167 evaluate the relative contributions of proximity to forest versus other factors on roadkill rates.
168 Mortality rates tended to be higher on straight stretches, perhaps because cars are travelling
169 faster.

170 Identification of factors that affect mortality rates of specific taxonomic groups from
171 hotspot maps can lead to efficient implementation of targeted mitigation strategies. Integrated
172 approaches, such as speed control and signage may help reduce wildlife road mortality
173 [27,28], although the success of road signs by themselves is limited [29,30]. Wildlife crossing
174 structures, such as culverts, underpasses, or overpasses may be particularly beneficial to
175 wildlife on roads through protected areas, where the higher diversity and abundance of
176 animals, including threatened and/or endemic species, may translate to elevated impacts on
177 biodiversity [31,32]. In a landscape already fragmented, like on the Ecuadorian coast,
178 reducing barriers to the movement of wildlife becomes an urgent necessity to prevent decline
179 and local extinction of animal populations [13,33,34].

180 We recommend culverts and underpasses be installed in hotspots of amphibian and
181 mammal mortality. These are particularly effective when combined with fences along the
182 road to guide the animals toward them [35]. Camera trapping has shown such structures are
183 used by wildlife, and are an effective means of reducing animal-vehicle collisions and
184 connecting wildlife habitat [36], although few “before” and “after” studies exist to allow
185 direct evaluation of impact [37]. We recommend the installation of signage and velocity
186 reduction devices (e.g., speed bumps) in straightaways with mortality hotspots. We hope that
187 our study will provide the needed baseline for monitoring if our recommendations for
188 mitigating wildlife mortality in coastal Ecuador are implemented.

189 **Geolocation information**

190 The northern end of the study area is the city of Pedernales (UTM 10007232 N, 0605361 E;
191 WGS84 zone 17S) and the southern end of the survey is the community of Rambuche
192 (9966571 S, 0574460 E). The approximate center is the Lalo Loor Dry Forest Reserve
193 (9991478 S, 0594199 E).

194 **Acknowledgements**

195 We are grateful to field assistants who assisted with data collection, including Stephanie
196 Bianco, Ryan Grist, Clea Harrelson, Michael Li, Marin Oschmann, Tomas Setubal, Kayla
197 Theis, Grace White, and Kylie Vanchena. We also thank the donors to the Ceiba Foundation
198 for Tropical Conservation who made this study possible, and the ESRI Nonprofit
199 Organization Program for reduced-cost software license.

200 **Author contribution**

201 Catherine L. Woodward co-designed the study, obtained funding, supervised field assistants,
202 conducted part of the surveys, and co-wrote the paper.

203

204 Luis M. Fernández performed data preparation, carried out GIS and statistical analyses, and co-wrote
205 the paper.

206

207 Joe E. Meisel co-designed the study, obtained funding, supervised field assistants, conducted some of
208 the surveys, carried out GIS analysis, and provided revisions on the paper.

209 **References**

210 [1] Lalo J. The problem of roadkill. *Am For.* 1987;50:50–52

- 211 [2] Forman RT, Sperling D, Bissonette JA, Clevenger AP, Cutshall CD, Dale VH, Fahrig L,
212 France R, Goldman CR, Heanue K, Jones JA, Swanson FJ, Turrentine T, Winter TC.
213 Road Ecology: Science and Solutions. Washington DC (CO): Island Press; 2003.
- 214 [3] Coffin AW. From roadkill to road ecology: A review of the ecological effects of roads. J
215 of Transp Geogr. 2007;15:396–406.
- 216 [4] Riley SJ. Effect of clearing and roading operations on the permeability of forest soils,
217 Karuah catchment, New South Wales, Australia. For Ecol and Manag. 1984;9(4):238–
218 293.
- 219 [5] Bager A, da Silva P, Bourscheit A, Kuczach A, Maia B. Os caminhos da conservação da
220 biodiversidade brasileira frente aos impactos da infraestrutura viária. Biodiversidade
221 Brasileira. 2016;6(1):75–86.
- 222 [6] Bennett AF. Roads, roadsides and wildlife conservation: a review. In: Saunders DA,
223 Hobbs RJ, editors. Nature Conservation 2: The Role of Corridors. Chipping Norton:
224 Surrey Beatty; 1991. p. 99–117.
- 225 [7] Jaeger JA, Bowman J, Brennan J, Fahrig L, Bert D, Bouchard J. Predicting when animal
226 populations are at risk from roads: an interactive model of road avoidance behavior.
227 Ecol Model. 2005;185(2-4):329–348.
- 228 [8] Fahrig L, Pedlar JH, Pope SE, Taylor PD, Wegner JF. Effect of road traffic on amphibian
229 density. Biol Conserv. 1995;73:177–182.
- 230 [9] Boarman WI, Sazaki M. 1996. Highway mortality in desert tortoises and small
231 vertebrates: success of barrier fences and culverts. In: Evink G, Zeigler D, Garrett P,
232 Berry J, editors. Transportation and wildlife: reducing wildlife mortality and
233 improving wildlife passageways across transportation corridors. Washington DC

234 (CO): U.S. Department of Transportation, Federal Highway Administration; 1996.
235 p.169–173.

236 [10] Ramsden DJ. 2003. Barn Owls and Major Roads: results and recommendations from a
237 15-year research project. Ashburton (UK): The Barn Owl Trust; 2003.

238 [11] Ferreras P, Aldama JJ, Beltrán JF, Delibes M. Rates and causes of mortality in a
239 fragmented population of Iberian Lynx *Felis pardina* Temminck, 1824. Biol Conserv.
240 1992;61(3):197–202.

241 [12] Mittermeier RA, Robles P, Mittermeier CG. Megadiversity. México City: CEMEX;
242 1997.

243 [13] Fahrig L. 2003. Effects of habitat fragmentation on biodiversity. Annu Rev of Ecol,
244 Evol, and Syst. 2003;34:487–515.

245 [14] Johnson WC, Collinge SK. Landscape effects on black-tailed prairie dog colonies. Biol
246 Conserv. 2004;115:487–497.

247 [15] Simmons JM, Sunnucks P, Taylor AC, van der Ree R. 2010. Beyond roadkill,
248 radiotracking, recapture and F_{ST} – A review of some genetic methods to improve
249 understanding of the influence of roads on wildlife. Ecol and Soc. 2010;15(1):9.

250 [16] Quintero-Ángel A, Osorio-Domínguez D, Vargas-Salinas F, Saavedra-Rodríguez CA.
251 2012. Roadkill rate of snakes in a disturbed landscape of Central Andes of Colombia.
252 Herpetol notes. 2012;5:99–105.

253 [17] Seijas AE, Araujo-Quintero A, Velásquez N. Mortalidad de vertebrados en la carretera
254 Guanare-Guanarito, estado Portuguesa, Venezuela. [Vertebrate mortality in the
255 Guanare-Guanarito road, Portuguesa State, Venezuela]. Rev de Biol Trop.
256 2013;61(4):1619–1636.

- 257 [18] Santander T, Freile JF, Loor-Vela S. Ecuador. In: Devenish C, Díaz-Fernández DF, Clay
258 RP, Davidson I, Yépez-Zabala, I, editors. Important Bird Areas Americas – Priority
259 sites for biodiversity conservation. Quito: Birdlife International; 2009. p. 187–196.
- 260 [19] Secretaría Nacional de Planificación y Desarrollo. Proyectos emblemáticos en Manabí.
261 [Emblematic projects in Manabí]. Montecristi: Secretaría de Planificación y
262 Desarrollo. Gobierno del Ecuador; 2003;p.88.
- 263 [20] Tirira DG. Libro rojo de los mamíferos del Ecuador. [Red book of Ecuadorian
264 mammals]. 2nd ed. Quito: Fundación Mamíferos y Conservación, Pontificia
265 Universidad Católica del Ecuador y Ministerio del Ambiente. Publicación especial
266 sobre los mamíferos del Ecuador 8; 2011.
- 267 [21] Birdlife International. The IUCN Red List of Threatened Species 2012:
268 e.T22678318A37858144. [http://dx.doi.org/10.2305/IUCN.UK.2012-](http://dx.doi.org/10.2305/IUCN.UK.2012-1.RLTS.T22678318A37858144.en)
269 [1.RLTS.T22678318A37858144.en](http://dx.doi.org/10.2305/IUCN.UK.2012-1.RLTS.T22678318A37858144.en). Downloaded on 12 September 2016.
- 270 [22] Puky M. Amphibian road kills: a global perspective. In: Irwin CL, Garrett P, McDermott
271 KP, editors. Proceedings of the 2005 International Conference on Ecology and
272 Transportation. Raleigh: Center from Transportation and the Environment, North
273 Carolina State University; 2006. p. 325–338.
- 274 [23] Glista DJ, DeVault TL, DeWoody JA. 2008. Vertebrate road mortality predominantly
275 impacts amphibians. *Herp Cons and Biol* 2008;3(1):77–87.
- 276 [24] D’Anunciação PE, Lucas PS, Silva VX, Bager A. Road ecology and Neotropical
277 amphibians: contributions for future studies. *Acta Herp* 2013;8(2):129–140.
- 278 [25] Antworth RL, Pike DA, Stevens EE. Hit and run: effects of scavenging on estimates of
279 roadkilled vertebrates. *Southeast Nat* 2005;4(4):647–656.

- 280 [26] González-Gallina A, Benítez-Badillo G, Rojas-Soto OR, Hidalgo-Mihart MG. The
281 small, the forgotten and the dead: highway impact on vertebrates and its implication
282 on mitigation strategies. *Biodivers Cons* 2013;22:325–342.
- 283 [27] Forman RT, Alexander LE. Roads and their major ecological effects. *Annu Rev of Eco*
284 *and Syst* 1998;29:207–231.
- 285 [28] Colino-Rabanal VJ, Lizana M, Peris SJ. 2011. Factor influencing wolf *Canis lupus*
286 roadkills in Northwest Spain. *Eur J of Wildl Res* 2011;57:399–409.
- 287 [29] Sullivan TL, Messmer TA. Perceptions of deer-vehicle collision management by state
288 wildlife agency and department of transportation administrators. *Wildl Soc Bull*
289 2003;31:163–173.
- 290 [30] Grilo C, Bissonette JA, Cramer PC. Mitigation measures to reduce impacts on
291 biodiversity. In: Jones SR. *Highways: Construction, Management, and Maintenance*.
292 Nova Science Publishers, Inc.; 2010. p. 73–114.
- 293 [31] Bager A. Repensando as medidas mitigadoras impostas aos empreendimentos viários
294 associados às unidades de conservação. [Rethinking the mitigation measures imposed
295 on road projects associated with the conservation units] In: Bager A., editor. *Áreas*
296 *Protegidas. Conservação no âmbito do Cone Sul*. [Protected Areas. Conservation in
297 the Southern Cone]. Pelotas: Editor's edition; 2003. p. 159–172.
- 298 [32] Bager A, Fontoura V. Evaluation of the effectiveness of a wildlife roadkill mitigation
299 system in wetland habitat. *Ecol. Eng.* 2013;53:31-38.
- 300 [33] Yanes M, Velasco J, Suarez F. Permeability of roads and railways to vertebrates: the
301 importance of culverts. *Biol Cons.* 1995;71(3):217–222.

- 302 [34] De Meester L, Declerck S, Stoks R, Louette G, Van de Meutter F, Bie TD, Michels E,
303 Brendonck L. Ponds and pools as model systems in conservation biology, ecology and
304 evolutionary biology. *Aquat Cons: Mar and Freshw00 Ecosyst.* 2005;15:715-725.
- 305 [35] Dodd CK, Barichivich WJ, Smith LL. 2004. Effectiveness of a barrier wall and culverts
306 in reducing wildlife mortality on a heavily traveled highway in Florida. *Biol Cons.*
307 2004;118:619-631.
- 308 [36] Donaldson BM. Use of highway underpasses by large mammals and other wildlife in
309 Virginia and factors influencing their effectiveness. In: *Proceedings of the*
310 *International Conference of Ecology and Transportation.* San Diego (CA); 2005. p.
311 433–441.
- 312 [37] Glista DJ, DeVault TL, DeWoody JA. 2009. A review of mitigation measures for
313 reducing wildlife mortality on roadways. *Landsc and Urban Plan.* 2009; 91:1–7.

314 Table 1. Total number of carcasses found, roadkill rates (individuals/kilometer/year), and
 315 annual estimates of kills per year (individuals) for each taxonomic group based on low (3
 316 days), medium (7 days), and high (15 days) carcass turnover rates along the 63-km stretch of
 317 road sampled.

| Taxa | Carcasses | | Roadkill rate (ind/km/year) | Annual estimate (individuals) |
|--------------|------------------|---------------|--|--|
| Amphibians | 501 | Low | 16.1 | 1015.9 |
| | | Medium | 34.6 | 2177.0 |
| | | High | 80.6 | 5079.6 |
| Reptiles | 103 | Low | 3.3 | 208.9 |
| | | Medium | 7.1 | 447.6 |
| | | High | 16.6 | 1044.3 |
| Birds | 225 | Low | 7.2 | 456.3 |
| | | Medium | 15.5 | 977.7 |
| | | High | 36.2 | 2281.3 |
| Mammals | 77 | Low | 2.5 | 156.1 |
| | | Medium | 5.3 | 334.6 |
| | | High | 12.4 | 780.7 |
| Total | 916 | Low | 29.5 | 1857.4 |
| | | Medium | 63.2 | 3980.2 |
| | | High | 147.4 | 9287.2 |

318

319 **Figures**

320 Figure 1. Study area. This satellite image shows the sampled 63 km length of the Ecuadorian
 321 coastal highway (E-15, highlighted in yellow) in the province of Manabí.

322 Figure 2. Percentage of carcasses (out of 916 total) found during roadkill surveys, by Class.

323 Figure 3. Hotspot maps of wildlife mortality for: A) All taxa, B) All minus amphibians, C)
 324 Amphibians, D) Mammals, E) Reptiles, and F) Birds.